

A Sequential Diffusion Method for ^{15}N Natural Abundance Measurement of Ammonium, Nitrate and Total Dissolved Nitrogen in Water Samples

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Abstract

Natural abundance of nitrogen-15 ($\delta^{15}\text{N}$) is a useful tool to estimate sources of nitrogen (N). The objective of the study was to clarify practical conditions of a sequential diffusion-based method for ^{15}N natural abundance of inorganic N and total dissolved N (TDN) with relatively high concentration. We tested the sequential diffusion method at 40°C for 24 hours for recovery of $\text{NH}_4\text{-N}$ and $\text{NO}_3\text{-N}$ (0–40 mg L^{-1}) and for that of TDN (0–4 mg L^{-1}). Results showed that the complete recovery was achieved for inorganic N with 0.3–30 mg L^{-1} and for TDN 0–3 mg L^{-1} . Furthermore, recovery rates for TDN declined when the amount of N exceeded 120 $\mu\text{g N}$. The time required for N recovery can be shortened to 24 hours by increasing temperature to 40°C. No discrimination of ^{15}N occurred during the whole process under the above conditions. In conclusion, the sequential diffusion method for ^{15}N natural abundance measurement can be applied to water samples including 0.2–20 mg L^{-1} for $\text{NH}_4\text{-N}$ or $\text{NO}_3\text{-N}$, and 0.25–3 mg L^{-1} for TDN. The volume for TDN recovery must be adjusted so that amount of TDN in solution is less than 90 $\mu\text{g N}$.

Key Words

PTFE, Devarda's alloy, MgO, agricultural area

Introduction

Nitrogen pollution of ground or surface waters is a serious matter of concern in the world. In agricultural areas, N in water exists in forms of $\text{NH}_4\text{-N}$, $\text{NO}_3\text{-N}$ and organic N, and the concentration is relatively higher than in forest areas or in sea water. Natural abundance of ^{15}N has been used to estimate sources of N in water or soil samples and to understand the occurrence of denitrification in agricultural land (Maeda et al., 2003; Maeda et al., 2011).

Solid samples are subjected to measurement of $\delta^{15}\text{N}$ using a mass spectrometer with an element analyzer. The diffusion method developed by Sorensen and Jensen (1991), uses a polytetrafluoroethylene (PTFE) tape that encloses a sulfuric-acid-pipetted filter to absorb NH_3 from the solution. Sorensen and Jensen (1991) proposed a sequential diffusion method for collecting ^{15}N -labeled NH_4 and NO_3 . Stark and Hart (1996) measured ^{15}N concentration of TDN in water samples for the first time by recovering $\text{NO}_3\text{-N}$ converted from TDN by the persulfate digestion. Holmes et al. (1998) and Sigman et al. (1997) applied the sequential diffusion method to $\text{NH}_4\text{-N}$ and $\text{NO}_3\text{-N}$ at natural abundance levels.

Previous studies on diffusion-based methods have some drawbacks. Because of a long period required for N recovery (3–14 day), N may be transformed to another by bacteria during the process. Holmes et al. (1998) proposed the temperature of 40°C to increase a N recovery rate but they did not consider the time for the N recovery. The accuracy of $\delta^{15}\text{N}$ measurement is reduced by contamination with inorganic N derived from the air-born NH_3 or impurities in chemical reagents (Sigman et al., 1997; Stephan and Kavanagh, 2009). Since the diffusion methods have been applied to relatively low N concentration samples (< 3 mg L^{-1}), the optimal ranges of higher N concentration remains uncertain.

In this study, we find out practical conditions for the sequential diffusion-based method to recover different forms of N as pretreatments of ^{15}N natural abundance measurement. We examined recovery rates of $\text{NH}_4\text{-N}$, $\text{NO}_3\text{-N}$ or TDN with relatively high concentrations at 40°C for 24 hours. The recovery time can be shorter if the recovery of N is complete under these conditions.

Materials and Methods

Manufacturing PTFE Traps

According to Holmes et al. (1998), a glass fiber filter (Whatman GF/D, ϕ 10 mm, combusted at 450°C for 4

hours) was placed on a piece of PTFE tape (Sigma-Aldrich, 25×50 mm). Twenty microliters of 2 M H₂SO₄ were pipetted onto the center of the GF/D filter, then the PTFE tape was immediately sealed with tweezers. The PTFE traps were kept in a high density polyethylene bottle to prevent NH₃ contamination from the atmosphere.

Recovery of Inorganic N

Recovery rates of NH₄-N and NO₃-N were examined using the PTFE trap (Sorensen and Jensen, 1991). The solution samples include known concentrations of NH₄Cl or KNO₃ (0–40 mg L⁻¹). In the subsequent procedure, NH₄-N recovery was followed by NO₃-N. The isotopic discrimination of each form of N during the process was also checked for 15 mg L⁻¹ samples. A volume of each sample for N recovery was determined so that the solution contained 40–200 µg N, which is required for the mass-spectrometry analysis.

Different volumes of samples (1–10 mL) with known concentrations of NH₄-N and NO₃-N, and NaCl (0.5 g in 10 mL, combusted at 450°C) was added into a 100 or 200 mL vial to increase the osmotic potential to prevent the PTFE traps from swelling. After addition of a PTFE trap and MgO (0.03 g in 10 mL, combusted at 450°C), the vials were immediately sealed with a butyl rubber cap, then shaken at 90 rpm at 40°C for 24 hours. After removing the PTFE trap from the solution, a new PTFE trap and DA (0.08 g in 10 mL) were added into the solution for NO₃-N recovery. The vials were again shaken at 90 rpm at 40°C for 24 hours. High pressure gas inside the vials was released to the air by inserting a needle into the rubber at 8 hour of the shaking. The PTFE trap that was taken out from the solution was rinsed with distilled water, and kept in a desiccator for 3 days until being dry.

Recovery of TDN

Recovery rates of TDN were tested using NH₄Cl and glycine. Ten milliliters of samples containing NH₄-N or glycine (0–4 mg L⁻¹) were mixed with 2 mL of the persulfate oxidation reagent (4 mg NaOH and 3 mg K₂S₂O₈ in 100 mL of deionized water) for the digestion to NO₃-N for 30 minutes at 121°C. The digested samples were transferred to 100 or 200 mL vials and NO₃-N derived from NH₄-N or glycine was collected in the same manner as the NO₃-N recovery except for the timing of gas release (at 30 min of the shaking). Because the persulfate oxidation reagent increases the osmotic potential, we tested for NH₄-N samples whether or not NaCl should be added to the samples as used for the inorganic N recovery procedure. The ¹⁵N discrimination during the whole process was checked for 3 mg N L⁻¹ of glycine samples. A volume of samples were calculated according to the amount of N (40–200 µg).

Chemical Analyses

Collected PTFE traps were subjected to measurements of N concentration and δ¹⁵N. The GF/D filter containing NH₄-N was taken out of the PTFE tape. The GF/D filter was added into distilled water in a polyethylene bottle, then shaken for 1 hour at 25°C. The extracts from all samples were analyzed for NH₄-N concentration using a continuous flow analyzer (Auto analyzer, QuAatro 2-HR, BLTEC). For the isotopic analysis, the GF/D filter was enclosed into a tin cup (φ 5 mm, 9 mm deep). The enclosed samples were analyzed with a mass spectrometer (DELTA V ADVANTAGE, Thermo Scientific) combined with an element analyzer (FLASH2000, Thermo scientific). This procedure was done within 24 hours because the tin cups were corroded with H₂SO₄. The original chemical reagents were analyzed for ¹⁵N to compare with those of recovered N in samples.

Results and Discussion

Recovery rates were nearly 100% for NH₄-N and NO₃-N samples including < 150 µg N (Figs. 1 and 2), which is corresponding to N concentrations < 30 mg L⁻¹. The results indicate that the recovering period can be shortened to 24 hours from 3–14 days (Sorensen and Jensen, 1991; Koba et al., 2010). On the other hand, recovery rates declined for samples including 150 µg N (N concentrations of 30 mg L⁻¹) for NH₄ and 210 µg (40 mg L⁻¹) for both inorganic N. Koba et al. (2010) reported that 4200 µg NH₄-N was completely recovered from 100 mL. The different results should be further investigated. In blank tests, NH₄-N was detected with 1.4–2.5 µg N (Fig. 1), which is consistent with Stephan and Kavanagh (2009). This contamination would be derived from the impurities in MgO or air-born NH₃. This level of NH₄-N contamination can be negligible when analyzing water samples with relatively high N concentration.

Recovery rates of TDN inputted as NH_4Cl were significantly improved by no addition of NaCl ($p < 0.05$, Fig. 3), when the $\text{NH}_4\text{-N}$ concentration was less than 3 mg L^{-1} . Even without NaCl addition, PTFE traps were not broken for all samples. Accordingly, NaCl should not be added to TDN samples.

Recovery rates for TDN were nearly 100% for samples with $< 90 \mu\text{g N}$ (Figs. 4–6). The largest standard deviations were found for 2 mg N L^{-1} samples (150 and $120 \mu\text{g N}$ for 80 and 60 mL , respectively, Figs. 4 and 5). Recovery rates of TDN derived from glycine declined with larger standard deviations when the amount of N exceeded $120 \mu\text{g}$ (Fig. 7), presumably due to the limited capacity of $\text{NH}_4\text{-N}$ absorption by the PTFE trap used. Koba et al. (2010) reported that $4200 \mu\text{g N}$ was collected in the same PTFE trap during the process of $\text{NH}_4\text{-N}$ recovery from freshwater and 2 M KCl solution. The persulfate reagent with pH 13 would neutralize the acid in the PTFE trap, resulting in lower $\text{NH}_4\text{-N}$ absorption by the trap.

There was no significant difference in $\delta^{15}\text{N}$ values between the original reagents and recovered N ($p > 0.05$, Table 1). These results agreed with previous studies (Koba et al., 2010; Sigman et al., 1997).

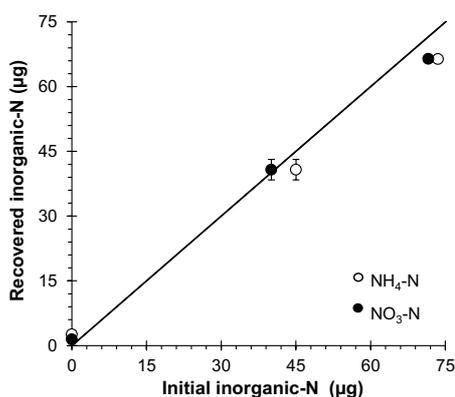


Figure 1 Initial vs. recovered inorganic N in low N concentration samples (Water volume: 150 mL).

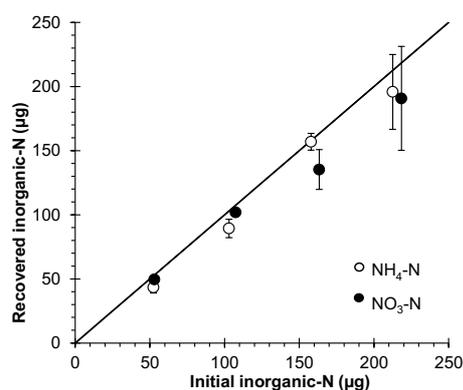


Figure 2 Initial vs. recovered inorganic N in high N concentration samples (Water volume: 5 mL).

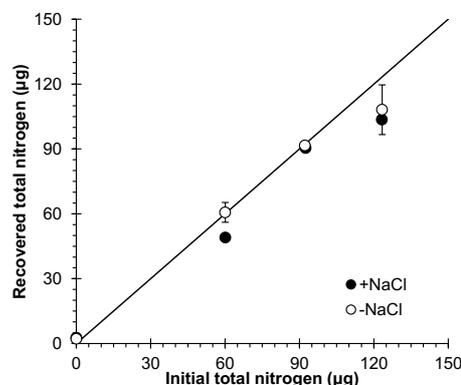


Figure 3 Effect of NaCl addition on the recovery process of TDN inputted as NH_4Cl (Water volume: 30 mL).

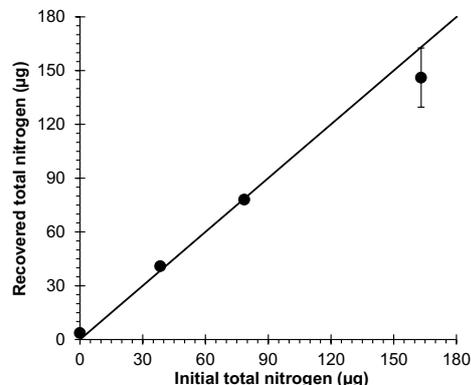


Figure 4 Initial vs. recovered TDN inputted as glycine with N concentration of $0.5\text{--}2 \text{ mg L}^{-1}$ (Water volume: 80 mL).

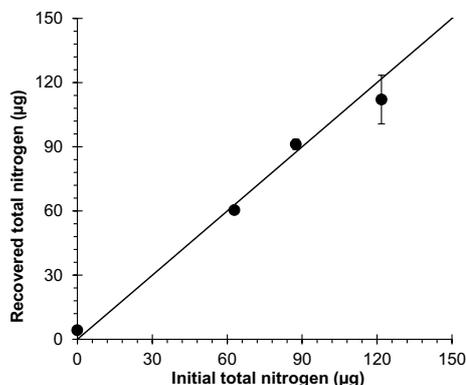


Figure 5 Initial vs. recovered TDN inputted as glycine with N concentration of $1\text{--}2 \text{ mg L}^{-1}$ (Water volume: 60 mL).

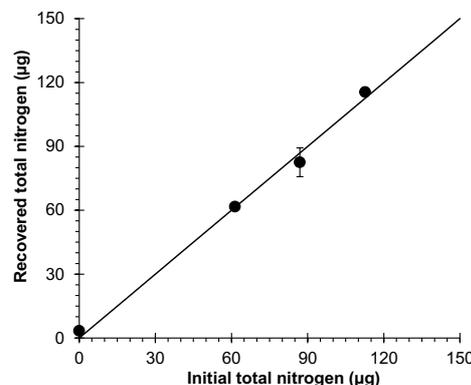


Figure 6 Initial vs. recovered TDN inputted as glycine with N concentration of $2\text{--}4 \text{ mg L}^{-1}$ (Water volume: 30 mL).

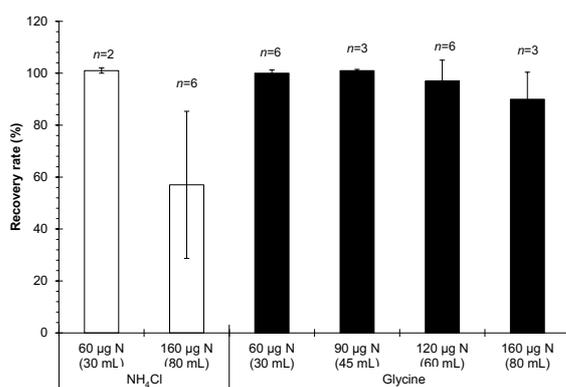


Figure 7 Effect of the amount of N on the recovery rates in NH₄Cl and glycine solutions with N concentration of 2 mg L⁻¹.

Table 1 Comparison of $\delta^{15}\text{N}$ values between the original reagents and recovered N.

Sample	N concentration (mg L ⁻¹)	$\delta^{15}\text{N}$ (‰)
NH₄-N		
Original NH ₄ Cl	–	-0.9±0.1*
Recovered N	15.0	-1.0±0.1*
NO₃-N		
Original KNO ₃	–	-1.8±0.0*
Recovered N	15.0	-1.9±0.2*
TDN		
Original Glycine	–	+0.1±0.0*
Recovered N	3.0	+0.2±0.1*

Note. * indicates that there is no significant difference in $\delta^{15}\text{N}$ values between the original reagent and recovered N. The analytical errors for $\delta^{15}\text{N}$ values are less than 0.2‰.

Conclusion

We examined practical conditions for the sequential diffusion-based method at 40 °C for 24 hours to recover N with different forms and high concentration. The time required for N recovery can be shortened to 24 hours by increasing temperature (40°C) and additionally by releasing the pressure inside the vial with a needle 1 hour after the DA addition. Each concentration of NH₄-N and NO₃-N should be 0.3–20 mg L⁻¹ and TDN 0.25–3 mg L⁻¹. For the $\delta^{15}\text{N}$ analysis of TDN, the amount of N should be less than 90 µg in a vial. If a water sample includes more than the ranges of N suggested above, the sample can be simply diluted before the N recovery procedures.

References

- Holmes R. M., McClelland J. M., Sigman D. M., Fry B., and Peterson B. J. (1998). Measuring ¹⁵N-NH₄⁺ in marine, estuarine and fresh waters: An adaptation of the ammonia diffusion method for sample with low ammonium concentrations, *Marine Chemistry* 60, 235-243
- Koba K., Inagaki K., Sasaki Y., Takebayashi Y., and Yoh M. (2010). Nitrogen isotopic analysis of dissolved inorganic and organic nitrogen in soil extracts, *Earth, Life, and Isotopic*, Kyoto University Press, 17-36
- Maeda M., Nakasone Y., Okamoto T., Asano Y., Fujiwara T., Nagare H., and Akao S. (2012). Reduced leaching losses of nutrients by using a catch crop during a fallow period following eggplant production in a green house, *Journal of Japanese Society of Civil Engineering G (Environmental Research)* 68, 103-111
- Maeda M., Zhao B., Ozaki Y., and Yoneyama T. (2003). Nitrate leaching in an andisol treated with different types of fertilizers, *Environmental Pollution*, 121, 477-487
- Sigman D. M., Altabet M. A., Michener R., McCorkle D. C., Fry B., and Holmes R. M.: Natural abundance-level measurement of the nitrogen isotopic composition of oceanic nitrate: an adaptation of the ammonia diffusion method, *Marine Chemistry* 57, 227-242, 1997
- Sørensen P. and Jensen E. S. (1991). Sequential diffusion of ammonium and nitrate from soil extracts to a polytetrafluoroethylene trap for ¹⁵N determination. *Analytica Chimica Acta*, 201-203, 1991
- Stark J. M. and Hart S. C. Diffusion technique for preparing salt solutions, Kjeldahl digests and persulfate digests for nitrogen-15 analysis, *Soil Science Society of America Journal* 60, 6, 1996
- Stephan K. and Kavanagh K. L. (2009). Suitability of the diffusion method for natural abundance nitrogen-15 analysis, *Soil Science Society of America Journal* 73, 2009